

# BAYESIAN TRACKING OF THE TOXIC PLUME SPREADING IN THE EARLY STAGE OF RADIATION ACCIDENT

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## KEYWORDS

Pollutant Spreading, Data Assimilation, State-Space models, Particle Filtering, Resampling

## ABSTRACT

The article deals with the predictions of time and space evolution of pollution dispersion during the early phase of a potential radiation accident. The goal is to design a proper fast algorithm which could enable more precise online estimation of radioactivity propagation on basis of recursive procedure of Bayesian filtering. Predicted trajectory of the plume of pollutants is refined online according to the values of observations incoming from terrain. The technique should be sufficiently robust to cope an expected lack of information in the same beginning of the event. A certain modification of the particle filter (PF) method is investigated here. Its robustness is illustrated on a real but atypical meteorological situation. Short time meteorological forecast entering the model is for this case in poor correspondence with the real time local meteorological measurements. Radiological measurements are assumed to be coming periodically from the Czech Early Warning Network (EWN). The respective radiological values in the real positions of EWN receptors are generated “artificially” drawing inspiration from the real local meteorological measurements.

## INTRODUCTION

Ongoing efforts on improvement of safety requirements cover both implementation of inherent safety features of the new constructed facilities and substantial improvement of emergency preparedness and response. Tracking and predictions of hazardous material spreading through the living environment provide decision-makers fundamental information for effective emergency management. Modelers should be capable to generate relevant information even in the lack of some basic input information. Correct chain of simulated consequences requires as realistic as possible description of the accident evolution from the same beginning of the harmful substances release. Just at the moment the accident scenario is not known completely and large uncertainties are involved. The evolution of emergency situation is usually so far varied and complicated that specific ad hoc solutions have to be introduced.

In this paper we are studying an application of data assimilation (DA) procedure insisting in optimum combination of prior knowledge with real observations incoming from terrain. The observations bring simultaneously an indirect information related to the system state. Advanced statistical

assimilation methods account for both model and measurements error covariance structure. The problem of pollution spreading in the atmosphere is described by nonlinear and generally non-Gaussian model. The attention is focused on Bayesian tracking of the toxic plume propagation over the terrain. It was shown (e.g. Ducet et al. 2001, Doucet 2008, Hoteit et al. 2008, Moradkhani 2008) that except simple problems the Bayesian inference in such complex systems is not analytically tractable. Consequently, the technique implemented here tries to solve a certain particular task of recursive Bayesian filter by Monte Carlo simulations. The objective of tracking is to refine recursively model predictions on basis of incoming measurements. Tracking in Bayesian approach concerns of recursive evaluation of the state posterior probability density function (pdf) evolution based on all available information. The article addresses the Bayesian tracking procedure from the same beginning of the complicated toxic plume spreading under (possibly) incomplete scenario description.

## PROBLEM FORMULATION

We restrict our attention to the stochastic state-space models

$$\begin{aligned}x_t &= g(x_{t-1}) + w_t \\ y_t &= h(x_t) + v_t\end{aligned}\tag{1}$$

in discrete time steps  $t=1, \dots, T$ . Here,  $x_t$  is N-dimensional vector unobserved internal quantities describing state of the model at time  $t$ , and  $y_t$  is M-dimensional vector of measurements obtained during the time step  $\langle t-1; t \rangle$ . Nonlinear vector functions  $g()$  and  $h()$  describe evolution of the state in time, and mapping of the state to measurements, respectively. Disturbance (noise) vectors  $w_t$  and  $v_t$  are considered to be independent realizations of random variables with zero mean and known variances,  $Q_t$  and  $R_t$ , respectively.

Formalization (1) is intuitively appealing for stationary additive disturbances (noises). However, it may be misleading when e.g. variance of the disturbance is state-dependent. Then, we consider a slightly more general version of (1)

$$\begin{aligned}x_t &\sim p(x_t | x_{t-1}) \\ y_t &\sim p(y_t | x_t)\end{aligned}\tag{2}$$

Where  $p(x_t|x_{t-1})$  denotes probability density function (pdf) of random variable  $x_t$  given realization of  $x_{t-1}$ . Model (1) arises as a special case (2) for choice  $p(x_t|x_{t-1})=N(g(x_{t-1}),Q_t)$  and  $p(y_t|x_t)=N(h(x_t),R_t)$ . The recursion starts at  $t=0$  for  $x_0 \sim p(x_0)$  which is known as *prior pdf*.

Model (2) enforces too strong restrictions: (i) realization of state variable  $x$  at time  $t$  depends only on values of  $x_{t-1}$ , and (ii) realization of the measurement  $y_t$  depends only on current realization of the state  $x_t$ . These assumptions may seem very restrictive, however, wide range of different models can be converted into the form (2) under appropriate choice of state variable  $x_t$ . For example, when initial conditions of the process or time-invariant parameters of the pdfs are not known, they are considered to be part of the state. In that case,  $x_t$  is sometimes called the augmented state, however, we will not make such distinction. In this paper,  $x_t$  denotes aggregation of all uncertainty in the model. Specific meaning of different parts of the state will be discussed later.

State-space formulation has been used in DA problem in the later stages of accident in post-emergency phases. Long term evolution of  $^{137}\text{Cs}$  deposited on terrain was predicted recursively (Hofman et al. 2008a) using Kalman filter technique, which is an optimal estimator for linear functions  $g()$  and  $h()$  and Gaussian *pdfs* in (2). But such linear model is insufficient for formulation of more complicated problems arising in the early phase of accident (Rojas-Palma 2005) and more general nonlinear dynamic model (2) is required. Bayesian approach to estimation of unknown quantities  $x_t$  is based on recursive evaluation of posterior density  $p(x_t|y_{1:t})$  using the Bayes rule:

$$p(x_t | y_{1:t}) \propto p(y_t | x_t) p(x_t | y_{1:t-1})$$

$$p(x_{t+1} | y_{1:t}) = \int p(x_{t+1} | x_t) p(x_t | y_{1:t}) dx_t. \quad (3)$$

Here,  $y_{1:t} = [y_1, \dots, y_t]$  and  $\propto$  denotes equality up to multiplicative constant, see (Ducet et al. 2001) for details. Note that since  $x_t$  aggregates all uncertainty in the model, posterior density  $p(x_t|y_{1:t})$  potentially provides estimates of unknown parameters, unknown initial conditions, or --- under appropriate parameterization --- even unknown variants of the model.

## PARTICLE FILTERING

### Principle

Except for few special cases (such as the Kalman filter), integration (3) is intractable. Therefore, various approximation has been proposed. The particle filter (also known as sequential Monte Carlo) is based on approximation of the posterior density by a weighted empirical approximation

$$p(x_t | y_{1:t}) \approx \sum_{i=1}^n w_{i,t} \delta(x_t - x_t^{(i)}) \quad (4)$$

where  $x_t^{(i)}$ ,  $i=1, \dots, n$  are samples of the random variable, i.e. the particles, and  $w_{i,t} > 0$ ,  $\sum_{i=1}^n w_{i,t} = 1$  are particle weights. Under this approximation, integration (3) is re-

duced to sampling from proposal densities (in our case  $p(x_t|x_{t-1})$ ), and recursive evaluation of particle weights  $w_{i,t}$ .

$$w_{i,t} \propto p(y_t | x_t^{(i)}) w_{i,t-1} \quad (5)$$

Key advantages of this approximation are easy evaluation of an arbitrary moment,  $m(x_t)$ ,

$$m(x_t) = \sum_{i=1}^n w_{i,t} m(x_t^{(i)}) \quad (6)$$

ability to handle arbitrary non-linear functions, and guaranteed convergence to the true posterior with growing number of particles  $n$ . The main disadvantage of the approach is its excessive computational cost.

### Adaptation of particle filtering scheme to the early phase of the plume propagation

Intuitively, the key state variable of the scenario is distribution of the pollutant in the atmosphere on the terrain. We model this distribution via segmented Gaussian plume model (SGPM). This is a discrete model with one-hour time step. Within each hour, given amount of a pollutant is released and evolution of this quantity is simulated taking into account all environmental effects (Pecha et al. 2007).

Real release dynamics is partitioned into equivalent number of fictive one-hour segments of constant release source strength. Synchronization with hourly forecast of meteorological conditions is performed. Hourly segment of the release is spreading during the first hour as a ‘‘Gaussian droplet’’. In the following hours of spreading according to available hourly meteorological forecast the droplet is treated as ‘‘prolonged puff’’ and its dispersion and depletion during the movement is simulated numerically by large number of elemental shifts. More detailed description of the procedure is described in (Pecha et al. 2008, Hofman et al. 2008). Each hourly segment  $g$  is consecutively modelled in its all hourly meteorological phases  $f$  and output vector  $v_{TOTAL}$  of values of interest are superposed as:

$$S_{TOTAL} = \sum_{(g)} \left\{ \sum_{f=1}^{F^{TOT}(g)} S^{g,f} \right\} \quad (7)$$

Each plume segment is uniquely described by the vector variable  $s^{g,f}$ . Evolution of each such plume over the terrain is described by deterministic SGPM model mentioned above. Let rewrite symbolically  $s^{g,f}$  to  $s(\tau)_t$ , where  $\tau < t$  denotes time of the release of the plume. The SGPM model contains many input and model parameters (Pecha et al. 2005). Most of them are treated as single values that enter the model by their best estimate values. Important random parameters are selected on basis of sensitivity analysis of the SGPM model and constitute random vector  $\Theta$ . Its dimension and meaning of the components selected for our scenario demonstrates Table 1.

Variable  $s(\tau)_t$  is now parameterized by vector of parameters  $\Theta_t$ . This vector contains both time invariant parameters, such are dispersion and dry deposition characteris-

tics, and time-variant parameters, such are wind direction and wind velocity at time  $t$ .

Under probabilistic formalization (2), the original SGPM model is interpreted as conditional density

$$p(s(\tau)_t | s(\tau)_{t-1}, \Theta_t) = \delta(s(\tau)_t - SGPM(s(\tau)_{t-1}, \Theta_t)) \quad (8)$$

Parameters  $\Theta_t$  were considered to be known in the original formulation. In this text, we consider them to be unknown, hence we consider them to be part of the state. The state is then  $x_t = [s(1)_t, s(2)_t, \dots, \Theta_t]$  and its evolution model

$$p(x_t | x_{t-1}) = \prod_{\tau=1}^t p(s(\tau)_t | s(\tau)_{t-1}, \Theta_t) p(\Theta_t) \quad (9)$$

Where distribution of vector parameter  $p(\Theta_t)$  is composed of independent pdf of scalar parameters given in table 1.

Table 1: Components of random parameter vector  $\Theta$ .

random parameter	unit	implementation inside code	uncertainty bounds
$\theta_1$ : act. release $f=1$	[Bq.h <sup>-1</sup> ]	$Q = c_1 \cdot Q^b$ $Q^b$ in $f=1$	LU; $c1 \in <0.31; 3.1>$
$\theta_2$ : horizont. dispersion	[m]	$\sigma_y(x) = c_2 \cdot \sigma_y(x)^b$	$N_{trunc}$ ; $c2 \in <0.89; 1.12>$
$\theta_3$ : dry depo velocity	[m.s <sup>-1</sup> ]	$vg = c_3 \cdot vg^b$	LU; $c3 \in <0.91; 1.10>$
$\theta_4$ : Wind direction $f=1$	[rad]	$\varphi = \varphi^b + \Delta\varphi$ , $\Delta\varphi = c_3 \cdot 2\pi/80$	U; $c4 \in <-12.0; +12>$
$\theta_5$ : Wind direction $f=2$	[rad]	$\varphi = \varphi^b + \Delta\varphi$ , $\Delta\varphi = c_3 \cdot 2\pi/80$	U; $c5 \in <-12.0; +12>$
$\theta_6$ : Wind direction $f=3$	[rad]	$\varphi = \varphi^b + \Delta\varphi$ , $\Delta\varphi = c_3 \cdot 2\pi/80$	U; $c6 \in <-12.0; +12>$
$\theta_7$ : Wind speed $f=1$	[m.s <sup>-1</sup> ]	$V_{10} = c_7 \cdot V_{10}^b$ $V_{10}^b$ in $f=1$	U; $c7 \in <0.5; 3.0>$
$\theta_8$ : Wind speed $f=2$	[m.s <sup>-1</sup> ]	$V_{10} = c_8 \cdot V_{10}^b$ $V_{10}^b$ in $f=2$	U; $c8 \in <0.5; 3.0>$
$\theta_9$ : Wind speed $f=3$	[m.s <sup>-1</sup> ]	$V_{10} = c_9 \cdot V_{10}^b$ $V_{10}^b$ in $f=3$	U; $c9 \in <0.5; 3.0>$
$\theta_{10}$ : act. release $f=2$	[Bq.h <sup>-1</sup> ]	$Q = c_{10} \cdot Q^b$ $Q^b$ in $f=2$	LU; $c10 \in <0.31; 3.1>$

Index  $b$  stands for "best estimate" values;

$V_{10}$  – wind speed at 10 m height;

$f$  – phase (hour) after the release start;

Type of distribution: LU-loguniform;  $N_{tr}$  – Normal, truncated;  
U – Uniform;

The measurements are modelled to have Gaussian distribution:

$$p(y_t | x_t) = N\left(\sum_{\tau=1}^t SGPM_{depo}(s(\tau)_t), \Sigma_t\right) \quad (10)$$

With mean value given by the sum of outputs from each plume, computed using model  $SGPM_{depo}$  which is a bilinear approximation of the SGPM model at the points of measurement. Covariance matrix  $\Sigma_t$  was chosen as

$$\Sigma_t = \sigma_{model} I_M + \sigma_{prop} \text{diag}(y_t) \quad (11)$$

with chosen constants  $\sigma_{model}$  and  $\sigma_{prop}$ . The first term models inaccuracies of the chosen Gaussian plume approximation,

the second term models inaccuracies of the measuring devices. This model is almost an arbitrary choice, that is used to show potential of the considered methodology. Model of observation for practical purpose should be designed using exact characteristics of the application specific measurement devices.

### Implementation of PF algorithm

The following steps represent computational flow of recursive particle filtering applied here:

1. Generate  $N$  realizations of parameter vector  $\Theta_0$  from densities listed in Table 1 and  $N$  corresponding plumes (in the following text interpreted as "particle"), assign all weight  $w_{i,t}=1/N$ .
2. For each time  $t=1..T$ 
  - a. Generate new realizations of  $\Theta_t$  and for each plume compute one step prediction using the SGPM. Let us introduce term "particle prolongation".
  - b. If measurements are available, recompute weights  $w_{i,t}$  using (5).
  - c. Compute posterior values of parameters of interest using (6)

Parameter vector  $\Theta$  is expressed in Equation (8) as  $\Theta_t$ . It means that count) of the components treated as random within a certain time interval can vary, symbolically:

$$\Theta_{t=1} \approx \theta_1 \theta_2 \theta_3 \theta_4 \theta_5 \theta_6 \theta_7 \theta_8 \theta_9 \theta_{10}$$

$$\Theta_{t=2} \approx \theta_1 \theta_2 \theta_3 \theta_4 \theta_5 \theta_6 \theta_7 \theta_8 \theta_9 \theta_{10}$$

$$\Theta_{t=3} \approx \theta_1 \theta_2 \theta_3 \theta_4 \theta_5 \theta_6 \theta_7 \theta_8 \theta_9 \theta_{10}$$

Relevant components entering the sample procedure are written in bold. We can also imagine alternative e.g. locally dependant land use characteristics when corresponding  $\theta_2$  and  $\theta_3$  could be assumed as relevant in all time steps.

### Experimental results

The sampling scheme consists of generation of 5000 particles corresponding to 5000 realisations of random parameter vector  $\Theta$  with 10 components  $\theta_i$  ( $i=1, \dots, 10$ ) according to uncertainty characteristics described in Table 1.

Evaluated values of the particle weights using  $\sigma_{model} = 100$  and  $\sigma_{prop} = cov \times e6$ , with  $cov=1, \dots, 5$ , are illustrated in Figure 1. The smallest values of variance (top) sharply selects only a few particles. With increasing variance,  $cov=2, \dots, 5$ , uncertainty in the weight grows and more particles become non-negligible.

Prior and posterior histograms of distributions of some parameters  $\theta$  from Table 1 are compared in Figure 2. Note that the posterior is sharply peaked for the three leftmost parameters while it is still widespread for the remaining parameters. But we should distinguish between parameter estimation for the concrete analysed situation and common average conditions. It should not be confused with parameter estimation which could give rec-

ommendation on parameter values commonly valid “in average”.

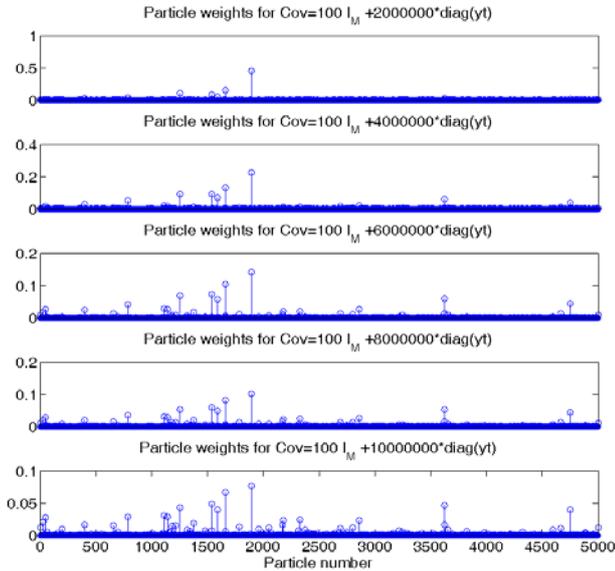


Figure 1: Posterior weights  $w_i$  at  $t = 2$  for five choices  $cov$ .

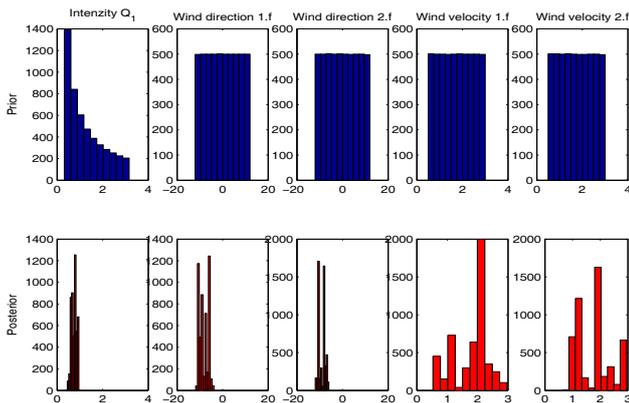


Figure 2: Comparison of prior (top row) and posterior (bottom row) histograms of distribution of selected parameters for  $cov=3$ .

**ILLUSTRATION OF PARTICLE FILTERING APPLIED IN THE EARLY STAGE OF A HYPOTHETICAL ACCIDENT**

The robustness of the PF method outlined above is illustrated for case of a certain circumstance when in the same beginning of an accident the decision maker is not provided by fully clear and unambiguous information. Experience from former radiation accidents pointed out the side effects leading to an information shocks with possible temporal paralysis of communication lines. In this sense we have adjusted a hypothetical accident scenario. Real meteorological situation from March 31, 2009 is taken into consideration and the moment of hypothetical accidental release is set to 10.00 UTC. Available real meteorological observations measured at the point of nuclear power plant (NPP) and short term meteorological forecast are somewhat inconsistent (see Table 2). Following ex post analysis can give a ret-

pective view on the atypical actual situations ( their occurrence rate is surprisingly not negligible). Due to a possible information shock mentioned above we shall assume conservatively a delay of two hours in recovery of radiation monitoring. Thus, the first measurements from terrain are coming just two hours after the release start. A decision maker has a dilemma how to manage the prediction of harmful substances in the early stage.

**Available meteorological data**

Let release of  $^{131}I$  radioactivity has started at 10.00 CET, March 31, 2009, and lasted for 2 hours (see Table 2). At this moment three kinds of meteorological data were directly available:

- Short term meteorological forecast generated twice a day (analysis time 00.00 and 12.00 UTM, for each hour, sequences up to 48 hours):
  - Label METLOC: Simple local forecast for the point of NPP (hourly sequences of wind direction and speed, category of atmospheric stability according to Pasquill and precipitation)
  - Label METGRID: extensive multilevel 3-D gridded meteorological forecast in HIRLAM format for vicinity  $160 \times 160$  kilometers around NPP
- Label METOBS: Observed values (real online meteorological measurements) incoming automatically from the point of NPP

All the data are provided by the Czech meteorological service and are available online through ORACLE DB server.

Table 2: Accidental release scenario of  $^{131}I$ , short-term meteorological forecast and real meteorological measurements for “point” of NPP Temelin (  $49^{\circ}10'48.53''N \times 14^{\circ}22'30.93''E$ ), time stamp 20090331-1000 CET.

CET hour	10.00	11.00	12.00	13.00
activity release of $^{131}I$ Bq/hour	$5.68 \times e+14$	$7.92 \times e+14$	0	0
wind direction <sup>1</sup> <b>METLOC/ME TOBS</b>	95.0 / 54.0	101.0 / 69.0	84.0 / 65.0	80.0 / 80.0
wind speed <sup>2</sup> <b>METLOC/ME TOBS</b>	2.0 / 3.8	2.1 / 3.0	1.9 / 3.8	2.2 / 3.8
Pasquill atm. stabil. <b>METLOC</b>	A	A	B	B

<sup>1)</sup> ... at 10 m height, blowing “from” (degrees measured clockwise from North); <sup>2)</sup> ... at 10 m height (m/s)

Deterministic calculations according to SGPM model with METLOC meteorology for the first two hours of the release are illustrated in Figure 3. Superposition according to (7) was used for quantity of  $^{131}I$  deposition on the ground ( $g=1, f=1$  and  $2; g=2, f=2$ ).

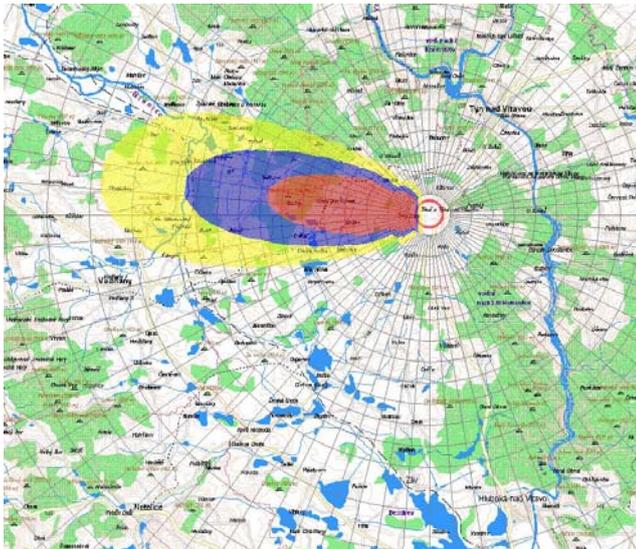


Figure 3: Release scenario with meteorological data METLOC - model predictions for “best estimate” values of model parameters, just 2 hours after the release start.

131I deposition ranges (Bq.m-2):  
 red:  $5.00e+06 \div 1.31e+08$  ; blue:  $1.00e+06 \div 5.00e+06$  ;  
 yellow:  $1.00e+05 \div 1.00e+06$  ;

**Arrangement of the real positions of monitoring sensors**

Early Warning Network (EWN) such a component of existing Radiation Monitoring Network (RMN) of the Czech Republic can be exploited for purposes of DA procedures. The main part of EWN is teledosimetric system (TDS) which for the NPP Temelin consists of two circles. The inner circle is positioned on the NPP-fence (see red circles in Figure 4 very close to NPP or in better discrimination in Figure 5) and consists from 24 stations 2,5m above ground. The outer II. circle measurement positions are drawn in Figure 4 by red squares. The dose-rate data are transferred each 4 minutes and stored to the ORACLE DB server for online access. We are assuming all these receptors to be operable. An ability to measure selected magnitudes of deposition is a question of a future monitoring development.

For DA purposes we have 79 sensors located in vicinity of the nuclear facility. In this number we have included 3 mobile measurement stations located randomly in the middle distances.

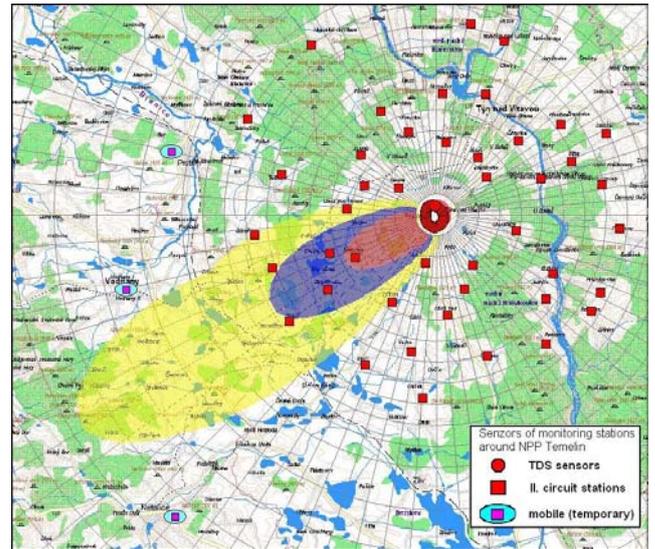


Figure 4: Release scenario with meteorological data METOBS - model predictions for “best estimate” values of model parameters, just 2 hours after the release start.

131I deposition ranges (Bq.m-2):  
 red:  $5.00e+06 \div 1.31e+08$  ; blue:  $1.00e+06 \div 5.00e+06$  ;  
 yellow:  $1.00e+05 \div 1.00e+06$  ;

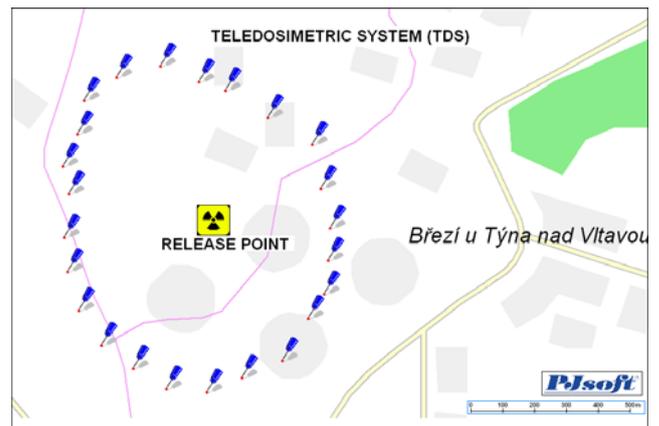


Figure 5: TDS on fence of NPP Temelin – 24 detectors

**Artificial simulation of the missing real accidental radiological data**

We hope that all considerations remain only in hypothetical level and the testing accidental radiological data will be always generated artificially. The technique is sometimes known as “twin experiment”.

A degree of belief to the initial near-range estimation using the SGPM model predictions with METLOC meteorological forecast (see Figure 3) will be low if we take into considerations the similar calculations with METOBS real meteorological measurement (see Figure 4). We should respect the fact that if something happens, the shape of the corresponding accidental trajectory close to the source will correspond more likely with the Figure 4. Without more discussion, we use this subjective assumption and generate the “artificial measurements” on basis of Figure 4.

## THE RESULTS ACHIEVED FOR SEVERAL FIRST TIME STEPS

Finally, the hypothetical DA scenario defined for the early phase includes:

1. Predictions of particles (trajectories) according to the SGPM model by given realisations of the parameter vector  $\Theta$ , always using gridded meteorological forecast METGRID.
2. The first set of “artificial measurements” (being in a certain discrepancy with model predictions) is incoming just two hours after the release start.
3. Update the posterior density according to measurements and evaluation of its selected moments.
4. Continuation of the recursive PF procedure in the next time intervals.

Generation of posterior density is performed for 5 choices of covariances  $cov=1, \dots, 5$  according to Equation (11) and numerical values from Figure 4. Expected mean values are calculated using common expression according to Equation (6), specifically in the form:

$$I(f_t) := E_{p(x_t|y_{1:t})} [f(x_t)] = \int f(x_t) p(x_t|y_{1:t}) dx_t \quad (12)$$

An estimation of the expectations on basis of  $N$  generated particles  $x_t^{(i)}$ ,  $i=1:N$  from posterior distribution is given by:

$$I^N(f_t) = \frac{1}{N} \sum_{i=1}^N f(x_t^{(i)}) \quad (13)$$

For  $N \rightarrow \infty$  is achieved almost sure convergence of  $I^N(f)$  to  $I(f)$ .

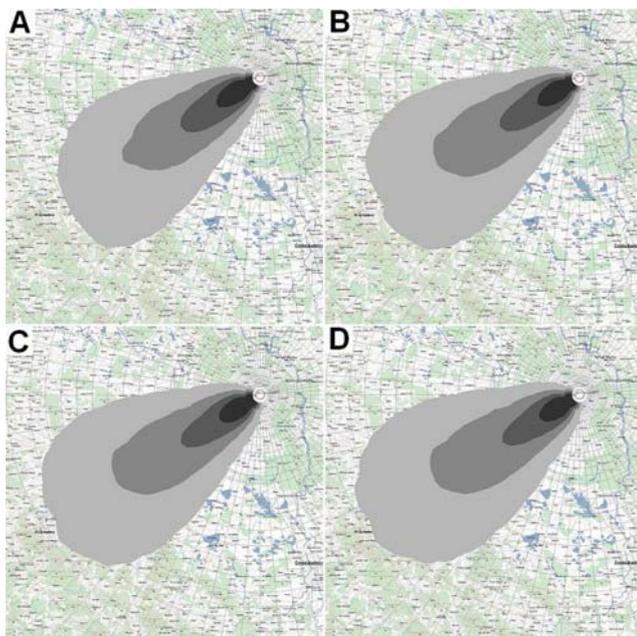


Figure 6: Expectations of the activity deposition in dependency on covariance matrix (acc. Equation (11)). A,B,C,D stand for  $cov=1,2,4,5$ .

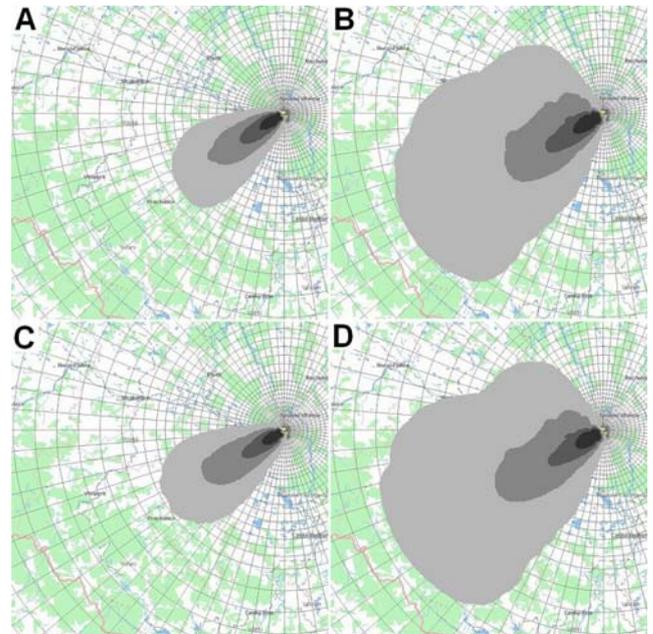


Figure 7: Prediction of expectations of the activity deposition quantity from 2. to 3. hour (case A  $\rightarrow$  B for  $cov=1$ ; case C  $\rightarrow$  D for  $cov=5$ ).

The expectations of the quantity of activity deposition are given in Figure 6 for cases of  $cov=1,2,4,5$ . The outer contour corresponds to the level of  $1.00 \text{ E}+03 \text{ Bq.m}^{-2}$ . The results show tendency of the updated model to approach the measurements with low noises. The values are slightly spreading when inaccuracies of measurements grows (higher  $cov$ ). Covariances of the measurement errors were selected rather low. At present new tests with increased covariance are running and tendency to lean to either model predictions or measurements are mapping.

Figure 7 demonstrates prolongation one time step forward. Case A concerns  $cov=1$  (also in Figure 6 A) expectation from the posterior density just after 2 hours after the release start. Using numerical approximation of the second part of Equation (3) which stands for transition equation and proposal  $pdf$  in specific formulation  $p(x_{f-2} | x_{f-3})$ , the prediction from the second hour (upper left A) to the third hour (upper right B) is done (prediction step). SGPM model prolongs the weighted particles within the step  $2 \rightarrow 3$ . The similar shift for  $cov=5$  stands for cases C  $\rightarrow$  D.

## CONCLUSION

The article extends former investigations in DA methodology (Hofman 2007) where analysis of the input model parameters uncertainty and both model error and observation error covariance structure were examined. DA in early stage of accident requires much more sophisticated access. From all possible techniques is adopted particle filter, which has one significant attribute. In PF the state ensemble trajectories are kept unchanged during the update step as for the forecast step and only their weights are updated. The particles remain unchanged after the correction (update) step and only receive the new weight (according to Equation (5)) reflecting closeness of the particle with respect the new observations.

This evident PF feature has favourable impact on exploitation of nonlinear prediction model SGPM in DA process in the early stage. SGPM model is in principle a trajectory model. The PF does not disrupt the trajectory information and it can be easily recursively forwarded.

The presented approach brings advantage of fast computation even for large number of realisations. One PF step of update and predictions with 5000 realisations is accomplished during 15 minutes and promises to support the decision making process in real time.

The adopted procedure seems to be robust and suitable to manage a certain discrepancies and scenario incompleteness occurring from the same beginning of accident. The authors narrow down anxiously the range of some uncertainties. For example the range of horizontal dispersion uncertainty  $c_2$  and dry deposition  $c_3$  should be much higher (in correspondence with expert judgments). Afterwards, the traces (e.g. in Figure 6) would be more dispersed in horizontal and longitudinal directions. Even the calculations have covered only the first time step and demonstrated code ability to predict in the second step, the full recursive PF application is easily feasible.

Still open remains a question of availability of measurements, capability to provide specific quantities and configuration and density of monitoring stations. The first negotiation between modellers and specialists responsible for monitoring was launched (Kuca 2008). The poor information can result from rare measurements. On the other hand, requirements issued from DA experience should be reflected in the future development of radiation monitoring networks.

DA plays substantial role in realistic prediction of evolution of radiation situation during nuclear emergency. Reliable information arriving on time provides decision makers with necessary time on judgement and introduction of efficient urgent countermeasures on population protection.

## ACKNOWLEDGEMENTS

This work is part of the grant project GAČR No. 102/07/1596, which is founded by Grant Agency of the Czech Republic. A lot of useful knowledge has been also acquired during RODOS customisation procedure for the Czech territory.

## REFERENCES

- Arulampalam, M. S.; S. Maskell, N. Gordon and T. Clapp. 2002. "A Tutorial on Particle Filters for Online Nonlinear/Non-Gaussian Bayesian Tracking". *IEEE Transactions on Signal Processing*, Vol. 50, No. 2.
- Drécourt, J-P. 2004. "Data assimilation in hydrological modeling". Environmental&Resources, Techn. Univ. of Denmark, ISBN 87-89220-84-6.
- Doucet, A.; N. De Freitas and N.J. Gordon (eds.). 2001. *Sequential Monte Carlo method in practice*. Springer-Verlag, New York.
- Doucet, A. and A.M. Johansen. 2009. "A Tutorial on Particle Filtering and Smoothing: Fifteen years later". *Will be published in Handbook of Nonlinear Filtering (2009)*.
- Moradkhani, H. 2008. "Hydrologic Remote Sensing and Land Surface Data Assimilation". *Sensors*, 8, 2986-3004, ISSN 1424-8220.
- Hofman, R. and P. Pecha 2007. "Integration of data assimilation subsystem into environmental model of harmful substances propagation". In *Proc. of the 11<sup>th</sup> Int. Conf. on Harmonisation within Atmospheric Dispersion Modelling* (Cambridge, UK, July 2-5, 2007), 111-115.
- Hofman, R.; P. Pecha ; and E. Pechová . 2008. "A simplified approach for solution of time update problem during toxic waste spreading in the atmosphere", *Hrvatski Meteorološki Časopis*, 12<sup>th</sup> Int. Conf. on Harmonisation within Atmospheric Dispersion Modelling (Cavtat, HR, Oct. 06-10, 2008)
- Hofman, R. and P. Pecha 2008a „Data Assimilation of Model Predictions of Long-Time Evolution of <sup>137</sup>Cs Deposition on Terrain“ , Proceeding of the IEEE International Geoscience & Remote Sensing Symposium 2008, IEEE International Geoscience & Remote Sensing Symposium 2008, (Boston, US, 06.07.2008-11.07.2008) (2008).
- Hoteit, I., D.-T. Pham, G. Triantafyllou and G. Korres. 2008. Particle Kalman Filtering for Data Assimilation in Meteorology and Oceanography. *Mon. Wea. Rev.*, 136.
- Kalnay, E. 2003. *Atmospheric modeling, data assimilation and predictability*. Cambridge University Press, ISBN 0-521-79179-0.
- Kuca, P., R. Hofman and P. Pecha. 2008. "Assimilation techniques in consequence assessment of accidental radioactivity releases – the way for increase of reliability of predictions". In *Proc. of ECORAD 2008 Int. Conf. on Radioecology & Environ.Radioactivity* (Bergen, Norway, June 15-20, 2008), 138-141, ISBN 978-82-90362-25-1.
- Pecha, P., R. Hofman and E. Pechova. 2007. "Training simulator for analysis of environmental consequences of accidental radioactivity releases". In *Proc. of the 6<sup>th</sup> EUROSIM Congress on Modelling and Simulation* (Ljubljana, Slovenia, Sept. 9-13, 2007), 18 pp. on conf. CD – ISBN 978-3-901608-32-2.
- Pecha, P. and E. Pechova. 2005. "Modeling of random activity concentration fields in air ... ". In *Proc. of the 10<sup>th</sup> Int. Conf. on Harmonisation within Atmospheric Dispersion Modelling* (Sissi, Greece, ), paper No. H11-069.
- Pecha, P. and R. Hofman. 2008. "Fitting of segmented Gaussian plume model predictions on measured data". In *Proc. of the 22th European Simulation and Modelling Conference ESM'2008*, (LeHavre, FR, 27.10.2008-29.10.2008) (2008)
- Rojas-Palma, C. 2005, Data assimilation for o\_site nuclear emergency management. Technical report, SCK-CEN, DAONEM final report, RODOS(RA5)-RE(04)-01, 2005.